

# Net Zero Teesside – Environmental Statement

Planning Inspectorate Reference: EN010103

# Volume III – Appendices Appendix 14D: Subtidal Benthic Ecology

**The Infrastructure Planning (Environmental Impact Assessment) Regulations 2017 (as amended)** 







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# **14D.Subtidal Benthic Ecology**

# **14.1 Introduction**

# **Project Background**

- 14.1.1 Net Zero Teesside Power Limited (NZT Power) and Net Zero North Sea Storage Limited (NZNS Storage), together the Applicants are seeking Development Consent for the construction, operation, maintenance and decommissioning of the Net Zero Teesside (NZT) Carbon Capture, Usage and Storage (CCUS) Project (the Proposed Development). The Proposed Development comprises the construction, operation and decommissioning of a CCUS facility comprising a gas-fired generating station with an electrical output of up to 860 MWe, together with equipment required for the capture and compression of carbon dioxide  $(CO<sub>2</sub>)$  emissions from the power generating station. In addition, there is a need for the provision of supporting infrastructure and connections to support the power generating station and to facilitate the development of a wider industrial carbon capture network on Teesside, the construction of which also forms part of the Proposed Development. The Proposed Development also includes high-pressure compression of  $CO<sub>2</sub>$  and the onshore section of a pipeline to export the captured  $CO<sub>2</sub>$  for off-shore storage.
- 14.1.2 The Proposed Development forms the onshore part of the wider NZT Project; further details relates to this are provided in Chapter 4: Proposed Development (ES Volume I, Document Ref 6.2).

# **Aims and Objectives**

- 14.1.3 The purpose of this report is to present the results of the subtidal benthic ecology surveys undertaken for this project, and to highlight key subtidal benthic receptors that may be affected by the development.
- 14.1.4 This report is intended to form part of the benthic ecological baseline characterisation study that will be undertaken to inform the various environmental assessments (e.g. Environmental Impact Assessment, Habitats Regulations Assessment, Water Framework Compliance Assessment) required to obtain development consent.
- 14.1.5 This report is not intended to formally characterise material for dredge and disposal purposes; the draft deemed Marine Licences includes a requirement for pre-construction sampling should dredging works be required.

# **Structure of Report**

- 14.1.6 This report is structured as follows:
	- **Section [14.2:](#page-4-1) Methodology** summarises the methodology for undertaking the subtidal benthic surveys as well as the approaches taken for sample and data analysis;





- **Section [14.3:](#page-14-1) Results** outlines the results of the subtidal benthic surveys;
- **Section [14.4:](#page-34-0) Discussion** discusses the results of the project-specific surveys in relation to existing publicly available information; and
- **Section [14.5:](#page-37-0) Summary of Findings** provides a summary of the findings of the project-specific surveys and a desk-based study for subtidal benthic ecology.

# <span id="page-4-1"></span>**14.2 Methodology**

14.2.1 The subtidal benthic ecology surveys were undertaken by Ocean Ecology Limited (OEL) on the 22 and 23 December 2019.

# **Study Area**

- 14.2.2 The Study Area was chosen by taking into the account the location of the Project and the predicted Zone of Influence (ZoI) of potential effects arising from the development. In addition, the Study Area was chosen in order to supplement the ground-truthing of existing information collected for Teesside Offshore Wind Farm (OWF). The Study Area encompasses and runs from Long Scar (7 km to the north) to Redcar Sands (7 km to the south) and up to 7.5 km offshore to the northeast.
- 14.2.3 The survey design was presented to and discussed with the Marine Management Organisation in September 2019; it included 23 sampling stations, from which triplicate sediment grab samples were collected (Figure 14D-1). Eight of these stations are located in the vicinity of the proposed Water Discharge Corridors / the Tees Bay, forming a 500 m x 500 m grid. A further three stations are situated within the main shipping channel within the Tees Estuary, along the south bank of the River Tees and adjacent to Paddy's Hole. In line with predominant tidal movements, two further far field stations (stations 3 and 4) were selected. Additional details of the sampling station locations, and rationale are provided in [Table 14D-1.](#page-4-0)



### <span id="page-4-0"></span>**Table 14D-1: Station locations and rationale**













*PSA = Particle Size Analysis; OWF = Offshore Wind Farm*



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<span id="page-7-0"></span>**Figure 14D-1: Subtidal benthic sampling stations** 



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**PROJECT** 

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 $63$ 

NET ZERO TEESSIDE PROJECT

Net Zero Teesside

#### **APPLICANTS**

NZT POWER LTD. AND NZNS STORAGE LTD.

<u>KEY</u>

Site Boundary

- Subtidal Study Area
- Mean Low Water
- Mean High Water

Subtidal Sampling Location - Sampling Method

- Macrobenthic/PSA
- Macrobenthic/PSA/Physico-Chemical

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FIGURE 14D-1 SUBTIDAL BENTHIC STATION LOCATIONS AND SAMPLING METHOD

**REFERENCE** NZT\_210511\_SBS\_14D-1\_v5



# **Survey Design**

- 14.2.4 The DSV Curtis Marshall survey vessel was used to collect and process the seabed sediment samples. A  $0.1$  m<sup>2</sup> Day grab was used to collect seabed samples, with three replicate samples per station. In total, 69 sediment samples (23 stations x 3 replicates) were taken for subsequent faunal and sediment particle size analysis. Additional replicate samples were also collected at 10 of the 23 stations for subsequent chemical analysis (see [Table 14D-1\)](#page-4-0).
- 14.2.5 Each retrieved grab sample was assessed for validity. Grab samples were deemed unacceptable and repeated if:
	- $\bullet$  the sample was less than 5 L;
	- the jaws of the Day grab failed to close;
	- the sample was taken from an unacceptable distance from the target location;
	- and there was obvious contamination of the sample from the equipment or debris.
- 14.2.6 If there were three failed attempts the station was moved 50 m away. This occurred for stations 3 and 9, where hard ground at the target locations resulted in multiple failed attempts.
- 14.2.7 Samples were processed aboard the survey vessel as follows:
	- 10% of sample removed for subsequent sediment particle size analysis (PSA) and transferred to labelled container;
	- the sediment sample was gently washed through a 1 mm sieve using a seawater hose; and
	- the remaining sample was backwashed into a container and preserved using a diluted formalin solution.

# **Laboratory and Data Analysis**

### Particle Size Analysis

- 14.2.8 Particle size analysis (PSA) was undertaken by OEL, on all 69 macrobenthic samples, undertaken in line with North East Atlantic Marine Biological Quality Control (NMBAQC) protocols (Mason, 2016), using dry sieving for the >1 mm fraction and laser diffraction for the fine fraction residue (<1 mm). Further information on analytical methods can be found in [Appendix A.](#page-43-0)
- 14.2.9 The dry sieve and laser data were merged for each sample with the results expressed as a percentage of the whole sample. Once the data was merged, PSA statistics and sediment classifications were generated from the percentages of the sediment determined for each sediment fraction using the Gradistat v8 software (Blott, 2010).
- 14.2.10 Sediment fractions were defined by size classes based on the Wentworth scale of particle size (Wentworth, 1922) [\(Table 14D-2\)](#page-10-0). Statistics such as mean and median grain size, sorting coefficient, skewness and bulk sediment classes (percentage silt, sand and gravel) were also derived in





accordance with the Folk classification system (a method of classifying sediment based on particle sizes defined by Wentworth) (Folk, 1954).

#### <span id="page-10-0"></span>**Table 14D-2: Wentworth scale of particle size for defining sediment type (Wentworth, 1922)**



### Sediment Chemistry Analysis

14.2.11 All chemical and metal analysis was undertaken by SOCOTEC UK Limited in accordance with MMO Marine Licensing Requirements (MMO, 2018). [Table 14D-3](#page-10-1) summarises the analytics.

# <span id="page-10-1"></span>**Table 14D-3: MMO marine sediment analysis carried out by SOCOTEC UK Ltd.**



<sup>1</sup> A modification of the Wentworth scale, using a logarithmic ( $\varphi = -\log_2 \frac{D}{D}$  $\frac{D}{D_0}$ ) transformation







*NDIR = Non-dispersive infra-red spectrophotometry; ICP-MS = Inductively coupled plasma mass spectrometry; DBT = Dibutyltin; TBT = Tributyltin; GC-MS = gas chromatography mass spectrometry; DTI = Doppler tissues imaging; EPA = Environmental Protection Agency; ICES = International Council for the Exploration of the Sea*

#### Macrofaunal Analysis

- 14.2.12 Macrobenthic analysis was undertaken by OEL in line with the NMBAQC Processing Requirement Protocol (PRP) (Worsfold and Hall, 2010).
- 14.2.13 All biota that had been retained on the 1 mm sieve were identified to species level, where possible, and enumerated by trained benthic taxonomists using the most up to date taxonomic literature and checked against existing reference collections and the World Register of Marine Species (WoRMS) for the latest taxonomic nomenclature. Colonial taxa (e.g. hydroids and bryozoans) were identified to species level where possible and recorded as present (P).
- 14.2.14 Major group biomass (Annelida, Crustacea, Mollusca, Echinodermata and Other taxa) was measured to the nearest 0.0001 g blotted wet weight. As a standard, conventional conversion factors, defined by Eleftheriou and Basford (1989), were then applied to provide equivalent dry weight biomass (Ash Free Dry Weight). The conversion factors applied were:
	- Annelida =  $15.5\%$ ;
	- Crustacea =  $22.5\%$ ;
	- Mollusca =  $8.5\%$ ;
	- Echinodermata = 8.0%; and
	- Other =  $15.5 \%$ .
- 14.2.15 A single reference collection, preserved in 70% IDA of all taxa identified, was retained for Quality Assurance (QA) purposes.
- 14.2.16 The macrofaunal community structure and diversity was analysed using the following parameters:
	- abundance (N);
	- species richness (S) (total number of species);
	- species diversity (H' loge) (Shannon-Wiener index<sup>2</sup>); and
	- biomass (g).



<sup>&</sup>lt;sup>2</sup> Shannon-Weiner index differs from species richness in that it takes into account not just the number of species but also the abundance of each species



#### Multivariate Analysis

- 14.2.17 The PRIMER v7 software package (Clarke and Gorley, 2015) was utilised to undertake multivariate statistical analysis on the macrobenthic dataset. In order to fully investigate the community assemblage patterns in the data, a suite of analytical routines was employed – the results for which can be found in Section [14.3.](#page-14-1)
- 14.2.18 To remove the weighting of common or rare species within a sample, the data was first transformed (in this instance square-root transformed<sup>3</sup>). A similarity matrix, which groups samples based on their community assemblage, was then constructed, for which the Bray-Curtis coefficient (S') was produced. Following this, cluster analysis was performed which provides 'natural groupings' of samples which is displayed in a dendrogram. An additional 'similarity profile' (SIMPROF) permutation test was then used to find statistically similar cluster groupings. In addition, multi-dimensional scaling (MDS) plots were created, which give a 2-dimensional representation of the similarity between samples. A Similarity Percentage analysis (SIMPER) test was then run to identify the individual species contributing the highest percentage to similarity within the cluster groupings.
- 14.2.19 To relate abiotic (environmental) factors to the biotic data, the BEST analysis was undertaken. This test identifies if any of the environmental variables measured are correlated with community patterns and provides a test statistic to determine which variables 'best explain'<sup>4</sup> the community patterns. The BIOENV routine completes this analysis in combination with determining a test of significance to give a variable/combination of variables which correlates highest to the biotic community data. Prior to the analysis, the sediment chemical data was log(x) transformed, and all abiotic data was normalised.
- 14.2.20 PRIMER v7 was also used to calculate a series of diversity metrics including species richness (S) and the Shannon-Weiner index (H' loge) at each station.

#### Ecological Quality

14.2.21 An assessment of the ecological quality of the infaunal communities was undertaken (Phillips *et al.,* 2014). The status of the sediment dwelling communities was determined using the Infaunal Quality Index (IQI) which was developed as part of the Water Framework Directive classification of transitional and coastal water bodies (Water Framework Directive TAG, 2008). The IQI is a multi-metric tool which utilises the AZTI Marine Biotic Index (AMBI), Simpson's evenness (a diversity index) and the number of taxa to produce an Ecological Quality Ratio (EQR) value which is a measure of the ecologically quality of infaunal communities. From this, an IQI status can be assigned to a benthic community which can range from bad (<0.24) to high (0.75). Where possible, the most up to date AMBI ecological group scores were used, provided via the AMBI software (http://ambi.azti.es). Granulometric data for each sample was also used to inform the IQI analysis,



<sup>&</sup>lt;sup>3</sup> Data was transformed in order to prevent skewness and the presence of outliers. This was indicated by using a Draftsman Plot (a version of a scatter plot).

<sup>4</sup> The variable or combination of variables with the highest correlation value. It must be noted that correlation does not necessarily imply causation.



as was the salinity regime ('coastal', 'mesohaline', 'oligohaline', 'polyhaline', and 'transitional').

- 14.2.22 The AMBI ecological groupings are as follows:
	- AMBI EG-I: "Species very sensitive to organic enrichment and present under polluted conditions (Initial state)";
	- AMBI EG-II: "Species indifferent to organic enrichment, always present in low densities with non-significant variations over time (from initial state to slight unbalanced)";
	- AMBI EG-III: "Species tolerant to excess organic matter enrichment. These species may occur under normal conditions, but their populations are stimulated by organic enrichment (slight unbalanced situations)";
	- AMBI EG-IV: "Second order opportunistic species (slight to pronounced unbalanced situations), mainly small sized polychaetes"; and
	- AMBI EG-V: "First-order opportunistic species (pronounced unbalanced situation)"

#### Habitat Classification

14.2.23 Environmental, sediment PSA and macrofaunal data obtained during the surveys was used to classify the habitats present in accordance with the European Union Nature Information System (EUNIS) classification system (EEA, 2012). An example is shown in [Table 14D-4.](#page-13-0) This classification system uses standard descriptions called 'biotopes', which categorise habitats based on the marine zone, the physical nature of the habitat and the biological communities observed. For example, marine habitats can be divided into littoral (also known as intertidal) and subtidal zones, and then classified according to the physical nature of the substratum, either rock or sediment, and the biological community found. Habitats observed were recorded to the lowest level possible.



### <span id="page-13-0"></span>**Table 14D-4: Example of the five-level EUNIS classification system (EEA, 2012)**







# <span id="page-14-1"></span>**14.3 Results**

# **Particle Size Distribution**

14.3.1 The major sediment fractions at each sampling station are presented in [Figure 14D-2.](#page-14-0) The PSA data has been summarised and classified as per the Folk (1954) classification system (as described in [Table 14D-5\)](#page-15-0). There was little variation between the stations located on the coast (excluding stations 1,2, and 5), most of them being dominated be sandy sediments, with a generally low mud (sediment <63  $\mu$ m) and gravel content (sediment  $\geq$ 2 mm). For the stations located in the mouth of the Tees estuary (stations 1, 2, and 5), mud represented the highest sediment fraction (>80%). The classification of most stations was 'slightly gravelly sand' (stations 3, 4, 7 - 21, 23), 'slightly gravelly muddy sand' (station 6), and 'sand' (station 22). The stations 1, 2, and 5, which had a higher content of mud, were classified as 'sandy mud'. See [Appendix B](#page-44-0) for full PSA results for replicate samples.



<span id="page-14-0"></span>**Figure 14D-2: Major sediment fractions (%) at each sampling station (replicate data averaged)** 





# <span id="page-15-0"></span>**Table 14D-5: Summarised PSA data as classified by Folk (1954) (replicate data averaged)**









# **Sediment Chemistry**

14.3.2 Sediment samples for contaminant analysis were collected at 10 of the 23 subtidal sampling stations (Table 14D-1: [Station locations and rationale\)](#page-4-0). Samples were analysed for heavy and trace metals, Polycyclic Aromatic Hydrocarbons (PAHs), Total Hydrocarbon Content (THC), Organotins, Polychlorinated Biphenyls (PCB) and Organochlorine concentrations. The full results of the sediment chemical analysis can be found in [Appendix C.](#page-48-0)

#### Heavy and Trace Metals

- 14.3.3 Concentrations of eight heavy and trace metals (arsenic, cadmium, chromium, copper, lead, mercury, nickel and zinc) were analysed for each of the 10 subtidal sampling stations. In the absence of any statutory thresholds, sediment concentrations have been compared to guidelines published by the Centre for Environment, Fisheries and Aquaculture Science (Cefas, 2003), and the Canadian Council of Ministers of the Environment (CCME, 1999) where applicable (i.e. no Cefas threshold available), to determine whether there is evidence of contamination.
- 14.3.4 The Cefas guidelines relate to the disposal of dredge material. There are two Cefas threshold levels; Action Level 1 (AL1) and Action Level 2 (AL2). In general, contaminant levels in dredged material which fall below (AL1) are of no concern. Levels above AL2 generally suggest that the dredged material is not suitable for sea disposal. Contaminant levels between AL1 and AL2 typically require further investigation.
- 14.3.5 The Canadian sediment quality guidelines consist of Threshold Effects Levels (TELs) and Probable Effects Level (PELs) (CCME, 1999). These thresholds have been derived from field research investigating the associations between chemicals and biological effects and the establishment of cause and effect relationships in certain marine organisms. At levels above the TEL, adverse effects may occasionally occur and at levels above the PEL, adverse effects may occur frequently (CCME, 1999).
- 14.3.6 Elevated levels of trace metals were recorded at stations 1 and 2, where the concentrations of all metals analysed, with the exception of cadmium, were greater than the Canadian guideline TEL. Concentrations of arsenic, chromium, nickel, lead, and zinc at stations 1 and 2 were also greater than the CEFAS guideline AL1. All other stations had relatively low levels of trace metals, with the exception of stations 10 and 11 for which concentrations of arsenic were greater than the Canadian guideline TEL. Of the trace metals recorded, no stations had concentrations which were greater than the respective CEFAS AL2 or the Canadian PEL thresholds.

### Hydrocarbon Concentrations (PAHs and THC)

- 14.3.7 Where available, PAH concentrations were compared to Effects Range Low (ERL) and Effects Range Medium (ERM) levels published by Long *et al*. (1995) as well as the Canadian TEL and PEL thresholds (CCME, 1999).
- 14.3.8 ERL and ERM concentrations are not thresholds of toxicity but delineate concentration ranges with associated probabilities of toxicity. Concentrations below the ERL represent a range in which detrimental effects on marine ecology would rarely be observed. Concentrations equal to or above the



ERL, but below the ERM, represent a range within which effects could be occasionally expected. Finally, concentrations equalling or exceeding the ERM represent a range within which effects could frequently be expected.

- 14.3.9 Similarly, Canadian TEL and PEL concentrations can be used as an assessment tool for identifying sediments in which adverse biological effects may occur (CCME, 1999). However, TELs and PELs should be treated as indicative, as they have been designed specifically for Canada and are based on the protection of pristine environments and species which may have different sensitivities to those in the North Sea.
- 14.3.10 Concentrations of PAHs were considerably higher at stations 1 and 2 compared to all other stations (see [Table C-2](#page-49-0) in [Appendix C\)](#page-48-0). At these stations, the concentrations of acenaphthylene, anthracene, benzo[a]anthracene, benzo[a]pyrene, diben[ah]anthracene, fluoranthene, and pyrene were all greater than the TELs and ERLs, whilst benzo[ghi]perylene and indeno[1,2,3-cd]pyrene were greater than the ERL only (this was the only standard available for these analytes). In addition to this, the concentrations of acenaphthene, fluorene, naphthalene, and phenanthrene were all greater than their respective PELs. The concentrations of naphthalene (station 1: 1,190 µm/kg, station 2: 1,410 µm/kg) and phenanthrene (station 1: 1,310 µm/kg, station 2: 1,440 µm/kg) in particular, were considerably higher than the PEL and close to the ERM, which was 2100 µm/kg and 1500 µm/kg, respectively. Furthermore, the concentration of chrysene at station 1 and 2 of 491 µm/kg and 515 µm/kg, respectively, was above the ERM (384 µm/kg). At all other stations, the concentrations of PAHs were below the standards presented, with the exception of naphthalene at station 11 which was above the TEL.
- 14.3.11 As with PAHs and trace metals, levels of THC at stations 1 and 2 were considerably higher than all other stations where concentrations were generally low. The United Kingdom Offshore Operators Association regards a value of 50 mg/kg to be the lower limit for a biological effect for THC (UKOOA, 2002). THC concentrations at stations 1 and 2, with concentrations of 581 and 334 respectively, were an order of magnitude higher than this.

### **Organotins**

14.3.12 Samples collected for contaminant analysis were analysed for the organotins: dibutyltin and tributyltin. All concentrations of organotins were found to be below the limit of detection of <0.001 mg/kg.

Polychlorinated Biphenyls (PCBs)

14.3.13 All total PCB concentrations sampled were below Cefas AL1 (0.01 mg/kg dry weight), except for station 2 which had a concentration of 0.011 mg/kg. In addition, all concentrations were below the Canadian TELs (21.5 mg/kg dry weight).

**Organochlorines** 

14.3.14 Organochlorines were compared to Cefas AL1 thresholds as well as OSPAR Background Concentration (BC) levels (OSPAR, 1998). BCs are assessment tools intended to represent the concentrations of certain hazardous substances that would be expected in the North-East Atlantic if





certain industrial developments had not happened. They represent the concentrations of those substances at "remote" sites, or in "pristine" conditions based on contemporary or historical data respectively, in the absence of significant mineralisation and/or oceanographic influences.

14.3.15 At the majority of stations, organochlorine concentrations fell below the limit of detection. This was with the exception of some organochlorine analytes at stations 1 and 2. However, concentrations of organochlorines at all stations remained below the OSPAR BC thresholds (0.050 mg/kg dry weight) and comparative Cefas AL thresholds.

**Macrobenthos** 

- 14.3.16 In total, 111 infaunal taxa were recorded across all 23 stations sampled. The macrobenthic community had a mean species richness of 14.4 (stdev =  $\pm$ 5.1; ranging from 2 to 28 per sample). The mean abundance across all samples was 787.5 individuals/m<sup>2</sup> (stdev =  $\pm$  504.5) whilst the average biomass was 6.12  $q/m^2$  (stdev =  $\pm$  12.98).
- 14.3.17 Annelida were the most abundant fauna recorded, representing 75.3% of the total average abundance across all the samples [\(Figure 14D-3\)](#page-19-0). Crustacea and Mollusca exhibited the second and third highest average abundances, measuring 87.2 individuals/m<sup>2</sup> (11.1%) and 86.8 individuals/m2 (11.1%), respectively. Echinodermata accounted for just 1.3% of the total average abundance, whilst the 'other' category contributed 1.3%. Phyla such as Platyhelminths and Nematoda, as well as phyla Nemertea and Phoronida were proportionally dominant within the 'other' group.
- 14.3.18 The average biomass across all samples, presented in [Figure 14D-3,](#page-19-0) showed a similar pattern to abundance with the phylum Annelida contributing the greatest proportion to average biomass (34.2%) followed by Crustacea (27.8%), Mollusca (25.1%) and Echinodermata (10.0%). 'Other' contributed the lowest to average biomass overall, representing  $0.2$  g/m<sup>2</sup> (3.0%).







# <span id="page-19-0"></span>**Figure 14D-3: Average abundance (individuals/m<sup>2</sup> ) (A) and biomass (g/m<sup>2</sup> ) (B) across all sampling stations for each phylum recorded**

- 14.3.19 Macrobenthic communities were found to be dominated by the polychaete *Magelona johnstoni* with this species representing 53.0% of the overall abundance recorded and an average abundance of 416.4 individuals/m<sup>2</sup> across the 23 stations sampled [\(Figure 14D-4\)](#page-20-0)*.* This species had an average abundance of 416.4 individuals/ $m^2$ , which represented the highest proportion of abundance overall (53.0%). All other taxa each contributed less than 10% to the total average abundance, demonstrating the low abundance but relatively high diversity of samples.
- 14.3.20 Two amphipod crustacean species were recorded in the top ten most abundant taxa including *Bathyporeia guilliamsoniana* and *Bathyporeia elegans*, with the former representing the second highest average abundance  $(52.5 \text{ individuals/m}^2)$  of all species sampled. The mollusc bivalve, *Fabulina fabula*, was the fourth most abundant species, with an average of 23.0 individuals/ $m^2$  recorded. The bivalve and polychaeta genus,





*Spisula* sp. and *Nephtys* sp. (respectively), are comprised of juvenile individuals which may explain their relatively high proportions of abundance compared to other taxa. [Appendix E](#page-58-0) presents the abundance of each taxon and biomass per major group (Annelida, Crustacea, Mollusca, Echinodermata and Others) in all samples collected across the survey area.



#### <span id="page-20-0"></span>**Figure 14D-4: Percentage (%) contribution of the ten highest recorded taxa to average abundance across all sampling stations**

14.3.21 The species richness (total number of species, S) and diversity (Shannon diversity index, H') at each station is presented in [Figure 14D-5.](#page-21-0) Species richness was highest at station 11 (S = 37), where diversity was also relatively high (H' = 2.303). The highest diversity of species was recorded at station 5 (H' = 2.854), which is located within the mouth of the Tees estuary. Stations 1 and 2 in this area also had values of diversity which were above average (H' = 1.962 and 2.305, respectively), although species richness recorded at station 1 had the fourth lowest value of species richness across all sites (S = 17). Species diversity was lowest at stations 22, 13 and 16, where values (H') of 1.296, 1.293, 1.275 was recorded, respectively. The lowest species richness was recorded at station  $9 (S = 11)$ , 13 (S = 11), and 17 (S = 8), which were all situated close to the shore [\(Figure 14D-5\)](#page-21-0).

### Priority Species and INNS

14.3.22 No species afforded conservation protection were recorded during the subtidal benthic grab surveys. Furthermore, no Invasive Non-Native Species (INNS) were recorded in any of the samples.







# <span id="page-21-0"></span>**Figure 14D-5: Average species richness (S) and Shannon diversity index (H') recorded at each subtidal station**

# **Multivariate Analysis**

- 14.3.23 [Figure 14D-6](#page-22-0) shows the non-metric MDS plot<sup>5</sup> of the community abundance data (square root transformed) from each sampling station, following the Bray-Curtis dissimilarity analysis. Samples with greater similarities in species composition are placed closer to one another on the plot with more dissimilar samples placed further away.
- 14.3.24 The MDS plot shows a clustering of the majority of samples, but with a high level of dissimilarity to the community composition of samples taken from stations 1, 2, and 5, which are located in the mouth of the Tees estuary. Some dissimilarity in faunal communities compared to other samples is also evident for stations 9, 13, and 17 and to a lesser extent, stations 8 and 18.
- 14.3.25 Cluster analysis and a SIMPROF test has identified 11 discrete groups of samples. The results of these tests have been overlaid on the MDS plot [\(Figure 14D-6\)](#page-22-0) and are reported in full on a cluster dendrogram in [Figure](#page-23-0)  [14D-7.](#page-23-0) These analyses show conformity between most stations (44 from a total of 69) assigned as group 'g', 'h', 'i' and 'f'. Group 'c' highlights the distinction of samples from stations 1, 2 and 5, although two replicates (1c and 5a) were not significantly similar ( $P < 0.05$ ), falling within group 'b' and 'e', respectively. Stations 13 and 17 have been segregated, falling within group 'd', whilst additional clustering is apparent of stations 8 and 18, assigned to group 'j', and all samples from station 6 which comprise group 'i'.



<sup>5</sup> A multi-dimensional scaling (MDS) plots gives a 2-dimensional representation of the similarity between samples.





<span id="page-22-0"></span>**Figure 14D-6: Non-metric MDS plot of community abundance data (square root transformed), with results of SIMPROF cluster analysis overlaid**

14.3.26 SIMPER analysis has identified the key taxa which contribute the most to within group similarity in community composition. The results of the analysis are presented in [Table 14D-6.](#page-24-0) The species *Magelona johnstoni* contributed the most to similarity in groups 'd', 'e', 'g', 'h', 'j' and 'k', with percentages ranging from 24.8% to 55.1%. *Bathyporeia* sp. were also dominant in four out of 11 groups, *Bathyporei guilliamsoniana* contributed highly to similarity in group 'h' and 'j'. The bivalve *Fabulina fabula* also was highlighted by the SIMPER analysis for groups 'h' and 'i'. Polychaete worms of the genus *Nephtys*, including *Nephtys cirrosa* and *Nephtys hombergii*, were of particular note in groups 'a', 'b', 'c', 'e', 'g', 'h', 'I' and 'j'.







<span id="page-23-0"></span>**Figure 14D-7: Cluster dendrogram of community abundance data (square root transformed), with results of SIMPROF cluster analysis overlaid**





## **Table 14D-6: Results of SIMPER analysis, comparing within cluster group similarity**



<span id="page-24-0"></span>*Nephtys assimilis* (6.92%)

*\*Less than two samples present in group and therefore no SIMPER results were produced. The species shown signify those that dominated the total abundance for that sample.* 





# **Environmental Data**

- 14.3.27 The non-metric MDS plot for sediment data, presented in [Figure 14D-8,](#page-25-0) shows the Folk (1954) classifications of each sample overlaid on the faunal MDS plot. This gives a description of the sediment characteristics at each station, providing some explanation as to the difference in the benthic species assemblage for each sample.
- 14.3.28 A clustering of samples classified as 'sandy mud' which were taken from stations 1, 2, and 5 (cluster groups 'b', 'c', and 'e') is evident. The substrate composition of the remaining samples was characterised by predominantly 'sand', with a small proportion of gravel or gravel with mud in some cases. These subtle differences in substrate composition did not appear to explain dissimilarity between the remaining samples. For example, samples belonging to cluster group 'd' represented the same substrate types (i.e. 'slightly gravelly sand' and 'sand') as the remaining cluster groups ('a', 'g', 'h', 'i' and 'f'), yet macrofaunal communities differed between these groups as indicated by the spatial separation of points shown on the faunal MDS plot.



#### <span id="page-25-0"></span>**Figure 14D-8: Non-metric MDS plot of community abundance data (square root transformed), with respect to Folk (1954) classification<sup>6</sup> for each sample**

14.3.29 A BEST analysis, using the BIOENV method, was undertaken to determine if various abiotic factors could explain the dissimilarity between samples. An initial analysis was run looking at both mean particle size, depth, and the percentage content of mud, sand, and gravel in the sediment of each sample. The results of the Global BEST Test indicated that a combination of both the percentage sediment content of mud and water depth provided the best explanation for faunal community dissimilarity ( $\rho$  = 0.679). The next best explanation for dissimilarity was depth and the percentage sediment content of sand ( $\rho$  = 0.678). Looking at each variable individually, the percentage



 $6$  S = 'sand', gS = 'gravelly sand', (g)S = 'slightly gravelly sand', (g)mS = 'slightly gravelly muddy sand', mS = 'muddy sand', sM = 'sandy mud',



sediment content of mud has the highest correlation value overall  $(p = 0.613)$ .

- 14.3.30 An additional BEST analysis was performed, testing a greater range of abiotic variables as follows: mean particle size, depth, percentage sediment content of mud, sand, and gravel, trace metals, and PAHs. Due to the methods of the sediment chemical analysis, the test could only be completed for stations 1, 2, 9, 10-15, and 17, for which an average per station was analysed. The Global BEST Test found that a combination of the PAHs (diben[ah]anthracene, phenanthrene, and pyrene), provided the greatest overall explanation ( $\rho = 0.892$ ) for the dissimilarity in species composition between stations. For each variable individually, the BEST analysis found c1-phenanthrene to have the highest value of correlation ( $\rho = 0.883$ ). Depth and the percentage sediment content of mud also provided some explanation to the variation between community assemblages ( $\rho = 0.874$ ) and  $\rho = 0.823$ , respectively).
- 14.3.31 Stations 1 and 2 exhibited elevated levels of PAHs compared to other stations but also had different sediment types and exposure regimes, being located in the mouth of the Tees Estuary. It should be noted that correlation does not always indicate causal effect, and it is likely a combination of variables is driving community dissimilarity.

# **Ecological Quality**

- 14.3.32 An EQR value has been calculated, providing an overall infaunal quality index (IQI) status using Water Framework Directive (WFD) classification metrics, which includes a combination of AZTI Marine Biotic Index (AMBI) scores (ecological groupings, AMBI-EG's I-V, of sensitive and opportunistic species), Simpson's evenness, and environmental data (sediment PSA and salinity) (WFD TAG, 2008; Phillips *et al.,* 2014). The results of these calculations are presented in [Figure 14D-9.](#page-27-0)
- 14.3.33 Almost all stations were categorised as having 'Good' IQI status. There were two stations with a High IQI status and two with 'Moderate' [\(Figure 14D-9\)](#page-27-0). Stations 2 and 5 which had a 'High' IQI status overall. These stations are located in the mouth of the Tees estuary and were characterised by the presence of *Nephtys hombergii, Nephtys* sp. and *Euchone sp.*, which are all assigned to AMBI-EG's II ('species indifferent'). Stations 13 and 17 which are located further inshore had an IQI status of 'Moderate'.







<span id="page-27-0"></span>**Figure 14D-9: Mean ecological quality ratio (EQR) scores (error bars represent standard error) at each station to inform the overall infaunal quality index (IQI) status**

# **Key Habitats and Species**

- 14.3.34 Across the subtidal benthic sampling stations, and based on the sediment PSA, three broad scale habitats were recorded. The stations which were coastal and dominated by a high content of sand, were classified as 'sublittoral sand' (A5.2). The stations in the mouth of the Tees Estuary, where the sediment content was high in mud, were classified as 'sublittoral mud' (A5.3). Variability in the species assemblage at each station and abiotic factors such as the composition of substrate, resulted in the ascription of three different biotopes:
	- '*Nephtys cirrosa* and *Bathyporeia* spp. in infralittoral sand' (A5.233; SS.SSa.IFiSa.NcirBat);
	- '*Fabulina fabula* and *Magelona mirabilis* with venerid bivalves and amphipods in infralittoral compacted fine muddy sand' (A5.242; SS.SSa.IMuSa.FfabMag); and
	- '*Nephtys hombergii* and *Macoma balthica* in infralittoral sandy mud' (A5.331; SS.SMu.ISaMu.NhomMac).
- 14.3.35 A description of these biotopes is presented in the section ['Biotope](#page-31-1) [Descriptions'](#page-31-1) below and a biotope summary table is provided in [Appendix D.](#page-53-0) A habitat classification map is presented in [Figure 14D-10.](#page-29-0)
- 14.3.36 Located in the mouth of the Tees Estuary, stations 1, 2, and 5 were classified as the biotope A5.331, falling within the biotope complex 'infralittoral sandy mud'. The sediment content in these areas were high in mud and supported relatively high abundances of the polychaete genus *Nephtys* sp., particularly *Nephtys hombergii* which is characteristic of this biotope. The bivalve *Abra alba* was also recorded, but in lower abundances. The biotope assigned to





these stations is typically associated with slightly reduced salinity estuarine conditions.

- 14.3.37 The second biotope recorded was A5.233, which is synonymous with sediment that has a high content of sand, with little to no fractions of mud ('infralittoral fine sand'). Stations 8, 9, 13, 17, and 18, which are located closest to the shore, were all classified as this biotope. This biotope is associated with sediments which are subject to higher levels of physical disturbance, as a result of wave action. The amphipod *Bathyporeia* sp. and polychaete *Nephtys cirrosa* are typical of this biotope and dominated the abundance of these stations. In particular, *Bathyporeia elegans* and *Bathyporeia guilliamsoniana* were found in high abundance. Magelonid polychaeta would be expected for this biotope, as demonstrated by the species *Magelona johnstoni* being recorded in high abundance at these stations.
- 14.3.38 The remaining stations were classified as the biotope complex 'infralittoral muddy sand', having dominant fractions of sand with a silt/clay component between 5% and 20%. Taking into consideration the community composition of these stations, the biotope A5.242 was ascribed. For example, stations 3 and 4, which were the furthest stations located to the southeast and northwest, respectively, were identified as this biotope, as were the stations located further offshore. This biotope is typically found in less exposed areas compared to the biotope A5.233, 'extending from the extreme lower shore down to more stable circalittoral zone at about 15-20 m' (EEA, 2019). Due to the higher content of mud for this biotope, a greater dominance of venerid bivalves is expected, as well as the bivalve species *Fabulina fabula* and the polychaete genus *Magelona* sp.. Both *Fabulina fabula and Magelona johnstoni* dominated the abundances at the stations classified as this biotope. Juvenile individuals of the bivalve genus *Spisula* were also recorded at some stations identified as this biotope.
- 14.3.39 Two of the biotopes identified (A5.233 and A5.242) qualify as habitats of principal importance being listed under Section 41 of the Natural Environment and Rural Communities (NERC) Act 2006 and belong to the habitat type, 'subtidal sands and gravels'. However, these habitats are not a qualifying feature of any nearby designated site.



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<span id="page-29-0"></span>**Figure 14D-10: Subtidal benthic EUNIS biotope and sediment classifications map**

A5.331 (SS.SMu.ISaMu.NhomMac) – 'Nephtys hombergii and Macoma balthica *in in fra li t t o ra l s a n dy m u d '*



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# **AECOM**

**PROJECT** 

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NET ZERO TEESSIDE PROJECT

Net Zero **Teesside** 

#### APPLICANTS

NZT POWER LTD. AND NZNS STORAGE LTD.

<u>KEY</u>

Site Boundary

Mean Low Water

Mean High Water

Subtidal Sampling Location

Folk Classification Biotope - Symbol Shape

- (g)S Slightly Gravelly Sand
- A (g)mS Slightly Gravelly Muddy Sand

■ S – Sano

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• sM – Sandy Muc

EUNIS Biotope - Symbol Colour

A5.233 (SS.SSa.IFiSa.NcirBat) –<br>Wephtys cirrosa and Bathyporeia spp. in in A5.233 (SS.SSa.IFiSa.NcirBat) – *infralittoral sand'* 

> A5.242 (SS.SSa.IMuSa.FfabMag) – 'Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand'

SHEET NUMBER 1 o f 1

T I T L E

FIGURE 14D-2 SUBTIDAL BENTHIC EUNIS BIOTOPE AND SEDIMENT CLASSIFICATION MAP

**REFERENCE** NZT\_210511\_SBS\_14D-2\_v6

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Salt Sca

High St

Redca



# **Biotope Descriptions**

<span id="page-31-1"></span>14.3.40 The following descriptions are based upon those outlined within the EUNIS habitat classification system (EEA, 2012). See [Appendix D](#page-53-0) for full summary table of station biotopes.

Infralittoral Fine Sand

A5.233 - Nephtys cirrosa and Bathyporeia spp. in infralittoral sand

MHCBI: SS.SSa.IFiSa.NcirBat

Stations: 8, 9, 13, 17, 18

Infaunal multivariate clusters: a, d, e, j, k

Depth Range: 0 – 30 m

Descriptions: Characterised by *Nephtys cirrosa* and *Bathyporeia* spp. (and sometimes *Pontocrates* spp.), found from the shallow sublittoral to at least 30 m depth. This biotope occurs within well-sorted medium and fine sands which are subject to physical disturbance, such as wave action. Compared to less disturbed biotopes, the faunal diversity is reduced, consisting of more actively-swimming amphipods.





# <span id="page-31-0"></span>**Figure 14D-11: Biotope A5.233 at station 17, sample C**

### Infralittoral Muddy Sand

A5.242 - Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand

MHCBI: SS.SSa.IMuSa.FfabMag

Stations: 3, 4, 6, 7, 10, 11, 12, 14, 15, 16, 19, 20, 21, 22, 23

Infaunal multivariate clusters: f, g, h, i, k

Depth Range: 0 – 20 m

Descriptions: Communities are dominated by venerid bivalves such as Chamelea gallina and may be characterised by a prevalence of Fabulina fabula and Magelona mirabilis or other species of Magelona (e.g. M. filiformis). Other taxa which are commonly recorded include: the amphipod Bathyporeia spp. and polychaetes such as Chaetozone





setosa, Spiophanes bombyx and Nephtys spp.. This biotope is typically found in stable, fine, compacted sands and slightly muddy sands in the infralittoral and littoral fringe.



# <span id="page-32-0"></span>**Figure 14D-12: Biotope A5.233 at station 11, sample C**

### Infralittoral Sandy Mud

A5.331 - Nephtys hombergii and Limecola balthica in infralittoral sandy mud

MHCBI: SS.SMu.ISaMu.NhomMac

Stations: 1, 2, 5

Infaunal multivariate clusters: b, c, e

Depth Range: 0 – 20 m

Descriptions: This biotope occurs in predominantly near-shore muds and sandy muds but can also found in mixed sediments. The substratum is typically rich in organic content and the community is often quite stable. The presence of the polychaete Nephtys hombergii and the bivalve Limecola balthica characterise this biotope. Other species which may be important include Abra alba and Nucula nitidoasa, although they may not necessarily occur at the same time or in high numbers. The taxa Spiophanes bombyx, Lagis koreni, and Echinocardium cordatum may also be present. In addition, this biotope can occur in estuaries where salinities may be slightly reduced and where Mya sp. may form a significant part of the community.





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<span id="page-33-0"></span>**Figure 14D-13: Biotope A5.233 at station 5, sample B**







# <span id="page-34-0"></span>**14.4 Discussion**

- 14.4.1 Three biotope complexes were recorded across the 21 sampling stations and these were found to represent three spatially discrete areas characterised by sediments of varying composition. Exposure is also likely to have been a contributing factor.
- 14.4.2 Stations sampled on the south bank of the River Tees within the mouth of the estuary were characterised by the biotope complex '*Nephtys hombergii* and *Macoma balthica* in infralittoral sandy mud'. Here conditions were found to be relatively sheltered with weak tidal streams (>1 knot) which enable the build-up of muds which providing optimum habitat for the taxa *Nephtys* sp., in particular *Nephtys hombergii*.
- 14.4.3 Sampling stations out in the Tees Bay were classified as either '*Nephtys cirrosa* and *Bathyporeia* spp. in infralittoral sand' or '*Fabulina fabula* and *Magelona mirabilis* with venerid bivalves and amphipods in infralittoral compacted fine muddy sand'. The former biotope complex was found in the shallow inshore area which is characterised by moderate to high exposure and sediments possessing a low clay/silt content. The latter biotope complex characterised stations which were located in most cases, in slightly deeper waters and were less exposed and exhibited a percentage of silt/clay.
- 14.4.4 Stations 6, 7, and 8 corresponded to those sampled in 2010 as part of a benthic survey undertaken for the Teesside OWF development (Entec UK Limited, 2011) and so the biotope classifications can be compared. Biotope classifications remained consistent at stations 7 and 8. However, at station 6 an increase in mud content within sediments had led to a shift in biotope from 'infralittoral mobile clean sand with sparse fauna' recorded in 2010 to '*Fabulina fabula* and *Magelona mirabilis* with venerid bivalves and amphipods in infralittoral compacted fine muddy sand'. Given the anticipated mobility of sediment in this area, as a result of the varying levels of exposure along this coast, this change is not unexpected.
- 14.4.5 The Proposed Development includes the option to construct a replacement Water Discharge Corridor into the eastern end of Coatham Sands bay, alongside the proposed  $CO<sub>2</sub>$  Export Pipeline. Further consideration has therefore been given to samples taken as part of the Teesside OWF development, some of which spatially overlap with sampling undertaken for the Proposed Development. A full review and summary of this data is presented in [Appendix F.](#page-65-0)
- 14.4.6 The data from 16 OWF grab samples (which encompass the area to the South East of Tees Bay) show that the biotopes at these stations are consistent with those found in the 2019 project subtidal survey. The biotopes found were either '*Nephtys cirrosa* and *Bathyporeia* spp. in infralittoral sand' (A5.233; SS.SSa.IFiSa.NcirBat) or '*Fabulina fabula* and *Magelona mirabilis* with venerid bivalves and amphipods in infralittoral compacted fine muddy sand' (A5.242; SS.SSa.IMuSa.FfabMag), distributed depending on water depth gradients and mud content. This data shows that these two biotopes are consistently distributed across the bay and that the benthic communities observed in 2019 are comparable to those observed in 2010 (details of the





analysis undertaken are provided below), indicating the infaunal communities across the bay are consistent spatially and temporarily.

- 14.4.7 The OWF benthic surveys recorded a number of individuals and colonies of *Sabellaria spinulosa.* This species forms biogenic reefs which is an Annex 1 habitat under the Habitats Directive, as well as being a priority habitat of principal importance being listed under Section 41 of the NERC Act 2006. The results of the OWF benthic surveys concluded that the abundance of *Sabellaria spinulosa* was not great enough to represent biogenic reef. No individuals of *Sabellaria spinulosa* were recorded at any of the subtidal benthic stations sampled in 2019.
- 14.4.8 Multivariate analysis determined that the grouping of the subtidal benthic communities within the Study Area can be explained in part by the sediment classification of the samples, in particular the percentage sediment content of mud. Physical environmental factors, such as general circulation, tidal currents and wave exposure play an important part in determining the local nature of sediments via the processes of siltation and erosion though biotic factors such as inhabiting species which stabilise the sediment are also important (Thrush, 1991).
- 14.4.9 The stations located on the coast are generally more exposed, where fine particulate matter is more likely to be winnowed away, accounting for the greater proportion of 'sand' in this area. The sediment particle distribution can determine the favourability of a particular environment to certain species (Dauvin *et al.,* 2004). Therefore, to a certain extent the differences seen in the faunal community at stations 1, 2, and 5 compared to other stations, can be attributed to the higher content of mud in this area. These muddy stations exhibited greater diversity than most other stations.
- 14.4.10 In contrast, stations 13, 17, and 18, which were located close to the shore at depths of <2.5 m, had sediment which contained no fractions of mud. In exposed shallow coastal waters, wave action, surge and storms in particular can be a source of natural physical disturbance to soft sediment macrobenthos species (Dolbeth *et al.,* 2009). At lower depths, changes in hydrodynamics and higher near-bed flow can influence food availability, sediment characteristics, sediment organic matter content, pore-water chemistry, microbial content, and larval supply (Incera *et al.*, 2003; Snelgrove and Butman, 1994). Higher levels of exposure are probable at these lower depth stations, apparent from the sediment classifications in these areas which were identified as being 'sand' and 'slightly gravelly sand'. This has likely altered the macrobenthos community in these areas compared to other stations.
- 14.4.11 Of the physical factors the percentage content of mud within the sediment, in combination with water depth, were the factors that best explained the differences in benthic community composition between stations. The gradients of abiotic factors such as light, water movement, nutrient availability, sedimentation and temperature can be predicted using depth (Garrabou *et al.*, 2002) and therefore depth is a variable which may be a proxy for many of these environmental conditions.
- 14.4.12 Sediment chemistry however, was also found to be an important factor the correlated with the nature of the benthic communities. There were elevated




levels of both trace metals and PAHs identified at stations 1 and 2, in the estuary with some above environmental thresholds. These elevated contaminants reflect the history and nature of the subtidal Study Area as a highly industrial region, with a broad variety of industries, including steelmaking and chemical manufacture, utilising land and resources within close proximity to the marine environment.

- 14.4.13 The concentrations of PAHs were compared against the sediment chemical analysis undertaken in 2015 to inform the PD Teesport Ltd maintenance dredging Marine Licence (MLP/2015/00094) (PD Teesport Ltd, 2015). At the sites analysed PAHs were also elevated (e.g. chrysene between 259 – 2,470 µg/kg) and encompassed the values recorded in 2019 at stations 1 and 2 (491 µg/kg and 515 µg/kg respectively). High levels of trace metals and PAHs can have toxic effects on infaunal communities, which can lead to long term changes, often reflected in the polychaete community assemblage (Papageorgiou *et al.*, 2006). The concentration of PAH, particularly the analyte c1-phenanthrene, was found to be the contaminant most correlated with distribution of the benthic communities.
- 14.4.14 Despite there being evidence of localised contamination, the ecological status of macrobenthic infaunal invertebrate assemblages at station 2 and 5 were both 'High', and at station 1 the status was 'Good'. Communities classified as 'High' are generally characterised by the presence of disturbance-sensitive taxa and levels of diversity and abundance associated with undisturbed conditions (Phillips *et al.*, 2014). Those assigned as having a 'Good' IQI status represents habitats which are only slightly disturbed. Both trace metals and PAHs adhere to sediment particles and are rapidly absorbed into the sediment following run-off into coastal waters (Dean, 2008). This reduces the bioavailability of these chemicals and their subsequent ingestion, which can provide a degree of protection to some infaunal taxa (Dean, 2008). Therefore, elevated sediment contaminant concentrations do not necessarily imply toxicity to benthic communities (Rees *et al.,* 2007) as the bioavailability of these chemicals is often more important than simply concentration levels.
- 14.4.15 Stations 13 and 17, identified as having both low species richness and diversity compared to other stations, had an ecological status of 'Moderate'. All other stations were found to be of 'Good' ecological status.
- 14.4.16 Two of the biotopes identified (A5.233 and A5.242) qualify as habitats of principal importance being listed under Section 41 of the NERC Act 2006 and belong to the habitat type, 'subtidal sands and gravels'. However, these habitats are not a qualifying feature of any nearby designated site.
- 14.4.17 No species afforded conservation protection were recorded during the subtidal benthic grab surveys. Furthermore, no INNS were recorded in any of the samples.





## **14.5 Baseline Evolution**

- 14.5.1 Benthic ecology baseline conditions can be influenced by a variety of factors including pollution, coastal development and climate change. These factors can influence not only the distribution of habitats and the abundance of associated flora and fauna but also life history processes such as growth and reproduction.
- 14.5.2 Within the Study Area, climate change impacts are likely to include factors such as warming sea temperatures, ocean acidification, sea-level rise, alterations in salinity and oceanographic patterns, and increased numbers of storms and marine heatwaves (Stocker, 2013). Sea temperature rise in particular, is considered to be the principle way in which subtidal benthic baseline conditions are likely to evolve during the life cycle of the Project and is therefore considered in further detail below.
- 14.5.3 Future UK Climate Projections 2018 (UKCP18) from the Met Office for the Stockton-on-Tees area (The Met Office, 2019), based on a 1981 – 2000 baseline<sup>7</sup>, uses a range of possible scenarios, classified as Representative Concentration Pathways (RCPs), to inform different future emission trends. RCP 8.5 has been used for the purposes of this assessment as a worst-case scenario.
- 14.5.4 Under RCP 8.5 a rise in global sea surface temperatures of 1.5°C by 2050 is predicted, increasing to a 3.2°C rise by 2100 relative to 1870 – 1899 temperatures. In UK waters, mean annual sea temperatures have risen by 0.8°C since 1870 and have continued to show consistent warming trends since the 1970s onwards (Genner *et al*., 2017). According to Lowe *et al*. (2009), the seas around the UK are projected to be 1.5 – 4  $\degree$ C warmer by 2100.
- 14.5.5 Increased sea temperatures have already had effects on marine communities in UK waters. In the North Sea, increasing temperatures and changes in pelagic primary production have resulted in the variation in community structure of benthic species (Moore and Smale, 2020). A northern distributional shift of species in the North Sea, as well some species moving into deeper waters, coinciding with increasing sea temperatures has also been observed (Hiddink *et al.,* 2015). Southern species which are expanding in extent have done so faster than those further north which have shown signs of retreat (Moore and Smale, 2020). An Ecological Niche Model has been produced for infauna species in the North Sea, based on a mean temperature increase of  $2.8$ °C, which has predicted future infaunal distribution shifts (Weinert *et al.*, 2016). Overall, 60% of the infaunal species analysed were predicted to shift towards the north with ranges of  $10 - 50$ km. Furthermore, a 72% reduction in suitable habitat was anticipated for subtidal infauna species.
- 14.5.6 However, the evidence of the effects of climate change on subtidal benthic communities is limited, most studies being restricted to a very specific area, comparing the difference between two time periods (Moore and Smale, 2020). Most species and habitats are subject to a range of drivers of change



 $7$  This baseline has been selected as it provides projections for 20-year time periods (e.g. 2020 – 2039).



and therefore determining the significance of one stressor such as increasing sea temperature is limited (Moore and Smale, 2020). Furthermore, given the location of the Proposed Development, subtidal benthic species found to be present within the Study Area are unlikely to be at the extent of their distributional range. As a result, it is not anticipated that distributional shifts would be easily observed. Therefore, it is currently difficult to predict what localised changes, if any, may occur within the vicinity of the Proposed Development as a result of increasing sea temperatures and climate change overall.

## **14.6 Summary of Findings**

- 14.6.1 The subtidal benthic Study Area can be divided into three biologically distinct areas, the sediment content, and the putative level of exposure, being important determining variables. The three areas are the stations in the mouth of the Tees Estuary, classified as the biotope complex 'infralittoral sandy mud'; the coastal stations located close to the shore, classified as 'infralittoral fine sand'; and the stations located further offshore or are far field, classified as 'infralittoral muddy sand'.
- 14.6.2 Three biotopes were recorded, those being: '*Fabulina fabula* and *Magelona mirabilis* with venerid bivalves and amphipods in infralittoral compacted fine muddy sand' (A5.242); '*Nephtys hombergii* and *Macoma balthica* in infralittoral sandy mud' (A5.331); and '*Nephtys cirrosa* and *Bathyporeia* spp. in infralittoral sand' (A5.233).
- 14.6.3 Two of the biotopes identified (A5.233 and A5.242) qualify as 'Habitats of Principle Importance' and 'Habitats of Conservation Interest'. However, the habitat types at each station were not a qualifying feature of any nearby designated site.
- 14.6.4 The subtidal Study Area is situated within a highly industrial region, with a broad variety of industries, including steelmaking and chemical manufacture, utilising land and resources within close proximity to the marine environment. Elevated levels of PAHs and trace metals were recorded at stations 1 and 2, and in some instances were greater than guideline concentration standards (Canadian guideline PEL and ERM defined by Long *et al.* (1995)).
- 14.6.5 The majority of stations were classified as having 'Good' IQI status, representing habitats that were only slightly disturbed. Stations located in the mouth of the Tees Estuary (2 and 5), were classified as having a 'High' IQI Status. In contrast, stations 13 and 17, located close to the shore where species richness and diversity was relatively low compared to other stations, had an IQI status of 'moderate'.
- 14.6.6 No species of any conservation designation or importance were recorded during the subtidal benthic grab surveys. Furthermore, no Invasive Non-Native Species (INNS) were recorded in any of the samples.
- 14.6.7 Prior to and during the construction and operational phase of the Proposed Development, the subtidal benthic baseline is likely to evolve as a result of climate change due to increases to both sea level and sea temperatures. This baseline evolution could result in a shift in the distribution of subtidal benthic species as well as a decrease in the availability of suitable habitat.





However, it is not possible to predict with any certainty the magnitude of potential changes to baseline conditions as a result of climate change or any other pressure.





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# **Appendix A : Particle Size Distribution (PSD) analysis methodologies**

## **Introduction**

The method used involved drying all sediments at 80°C for at least 24 hours prior to dry-sieving all samples and only laser sizing the <2 mm fraction if >5 % of the whole sample was found to be <63 μm. Oven drying sediment causes the aggregation of particles in muddy sediments (>5 % mud) and for these reasons, such sediments should not be oven dried prior to particle size analysis (Mason, 2016). Therefore, a visual assessment of all thawed sediment samples was undertaken prior to drying to ensure the optimal analysis technique was used. Due to the obvious presence of mud in a large proportion of samples, some with a considerable mud content in excess of 5 %, all samples were analysed via a combination of both dry sieving (>1 mm fraction) and laser sizing (<1 mm fraction).

### **Sample Preparation**

Frozen sediment samples were first transferred to a drying oven and thawed at 80°C for at least six hours prior to visual assessment of sediment type and wet sieving over a 1 mm sieve. Before any further processing (e.g. sieving or sub-sample removal), samples were mixed thoroughly with a spatula and all conspicuous fauna (>1 mm) which appeared to have been alive at the time of sampling were removed from the sample.

### **Dry Sieving**

The >1 mm fraction was then returned to a drying oven and dried at 80°C for at least 24 hours prior to dry sieving. Once dry, the sediment sample was run through a series of Endecott BS 410 test sieves (nested at 0.5 φ intervals) using a Retsch AS200 sieve shaker to fractionate the samples into particle size classes. The dry sieve mesh apertures used are given in Table A-1.

### **Table A-1: Sieve series employed for Particle Size Distribution (PSD) analysis by dry sieving (mesh size in mm)**

**Sieve aperture (mm)**



The sample was transferred onto the coarsest sieve at the top of the sieve stack, which was then shaken for a standardised period of 20 minutes. The sieve stack was then checked to ensure the components of the sample had been fractioned as far down the sieve stack as their diameter would allow. A further 10 minutes of shaking was undertaken if there was evidence that particles had not been properly sorted (e.g. veneers of silt overlying coarse fractions).

### **Laser Diffraction**

The fine fraction residue (<1 mm sediments) was transferred to a suitable container and allowed to settle for 24 hours before excess water was syphoned from above the sediment surface. The fine fraction was analysed by laser diffraction using a wet element Beckman Coulter LS 13320. Due to the silty nature of the sediments, ultrasound was used to agitate particles and prevent aggregation of fines.





## **Appendix B :Particle Size Distribution (PSD) Data**

### **Table B-1: Summarised PSD data for all stations and replicate data, as classified by Folk (1954)**





















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# **Appendix C : Sediment Chemical Analysis Results**

**Table C-1: Trace and heavy metal sediment concentrations against Cefas (2003) and Canadian guidelines (CCME, 1999)**







### **Table C-2: PAH sediment concentrations against Canadian guidelines (CCME, 1999) and ERLs/ ERMs (Long** *et al***., 1995)**















#### **Table C-3: Organotin sediment concentrations against Cefas (2003) standards**



*Values exceeding sediment concentrations highlighted in corresponding colour* 

#### **Table C-4: PCB sediment concentrations against Cefas (2003) and Canadian guidelines (CCME 1999)**







### **Table C-5: Organochlorine sediment concentrations against Cefas (2003) standards and OSPAR BCs (OSPAR, 1998)**







# **Appendix D : Sample Biotope Summary Table**



#### **Table D- 1: Biotope summary table for each station, ordered by cluster group**

**<sup>8</sup>** \*Species taken from results of SIMPER analysis. Where less than two samples present the species shown signify those that dominated the total abundance for that sample.

 $9$  A5.2 = 'Sublittoral sand'; A5.3 = 'Sublittoral mud'.

<sup>10</sup> A5.233 = 'Nephtys cirrosa and Bathyporeia spp. in infralittoral sand'; A5.242 = 'Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand'; A5.331 = '*Nephtys hombergii* and *Macoma balthica* in infralittoral sandy mud'.































# **Appendix E : Macrofauna Data**

**Table E-1: Average macrofauna abundance (individuals/m<sup>2</sup> ) per station. 'P' denotes presence only**

<b>Station</b>																	13			16					20			22	23
Annelida																													
Amphictene auricoma	$\sim$					$3 -$			10		$17 -$																	$3 -$	
Aonides oxycephala																		$3 -$											
Caulleriella alata																		$3 -$											
Chaetozone christiei	$\sim$				3	$7 -$					$37 -$	$\blacksquare$		30		43	$13 -$		27	$7 -$			$\sim$		33	3		3	20 20
Chaetozone gibber		$\overline{7}$		$10 -$				47 -																					
Chaetozone vivipara		$3 -$						$3 -$																					
Cirriformia tentaculata																		40 -											
Eteone longa		30	20		3	10	13		3	3		$10 -$			3	$\overline{7}$	$3 -$			$3 -$						3		$3 -$	$\overline{7}$
Euchone limnicola		140		243 -				$70 -$																					
Galathowenia oculata				$3 -$																									
Glycera lapidum	$\sim$												$7 -$				3	$7 -$											
Glycinde nordmanni						$3 -$					$7 -$				3	$23 -$	$\sim$		20	$7 -$					$3 -$			$13 -$	
Goniada maculata	$\sim$													$\sim$		$3 -$													
Lagis koreni	$\sim$								$10 -$		$\sim$				3	$3 -$													
Magelona filiformis		$3 -$			$10 -$				10	20		$3 -$		57		67	$23 -$		23	20	$7 -$		3		10	$\overline{7}$		$7 -$	17
Magelona johnstoni		10	13		287	523	30		250	437		180	47	877		623	950	20	610	583	657	37	93	573		797	660	453	867
Malacoceros jirkovi																		$13 -$											
Melinna palmata								$13 -$																					



































#### **Table E-2: Average macrofauna biomass (grams/m<sup>2</sup> ) per station**







## **Appendix F : OWF Data Analysis and Discussion**

### **Introduction**

Following the submission (to the Planning Inspectorate) of a Preliminary Environmental Information (PEI) Report in July 2020, for the Net Zero Teesside (NZT) project, a potential design change related to the Water Connection Corridor was identified. Specifically, this includes the potential relocation of the Water Discharge Corridor to a location to the south east of the current proposal, by constructing a new outfall in close proximity to the proposed CO2 Export Pipeline.

Should the new discharge infrastructure be required the Zone of Influence (ZoI) will move further east and therefore consideration of this new area has been addressed. In particular, a review of data in relation to the subtidal benthic baseline survey undertaken in 2019 was carried out because the 2019 subtidal survey was defined on the basis of the existing Water Discharge Corridor. The extension to the existing red line boundary, that would result from a new discharge corridor, is shown in Figure 14D-14.

The 2019 survey sampling stations did not extend into the newly proposed and amended red line boundary encompassing the newly proposed outfall. There are, however, several stations from the 2010 Teesside OWF benthic grab survey that do encompass this area (see Figure 14D-14, below) and provide subtidal community data within and in close proximity to the new outfall area.





**Figure 14D-14: Teesside OWF and Teesside Net Zero subtidal benthic sampling stations and EUNIS biotope classifications**

A5.331 (SS.SMu.ISaMu.NhomMac) – 'Nephtys hombergii and Macoma balthica *in in fra li t t o ra l s a n dy m u d '*

Phase I Intertidal Study Area

Intertidal Phase I Broad Scale EUNIS Biotope

A1.113, [Semibalanus balanoides] or exposed to moderately exposed or vertical sheltered eulittoral rock

A1.323, [Fucus vesiculosus] on variable salinity mid eulittoral boulders and stable mixed substrata

A2.2, Littoral sand and muddy sand

A2.242, [Cerastoderma edule] and polychaetes in littoral muddy sand

\* Entec UK Limited (2011). Teesside Windfarm Ltd, Teesside Offshore Wind Farm FEPA Monitoring, Benthic Survey Report 2010

T I T L E FIGURE 14D-14 TEESSIDE OFFSHORE WIND FARM AND NET ZERO TEESSIDE SUBTIDAL BENTHIC

EUNIS BIOTOPE AND SEDIMENT CLASSIFICATION MAP

**REFERENCE** NZT\_210511\_SBS\_14D-141\_v2



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## **AECOM**

**PROJECT** 

 $\mathbf{\hat{\boldsymbol{\theta}}}$ 

NET ZERO TEESSIDE PROJECT

Net Zero **Teesside** 

#### APPLICANTS

### NZT POWER LTD. AND NZNS STORAGE LTD.

Site Boundary <u>KE`</u>

PEIR Site Boundary

Mean Low Water

- Mean High Water
- **e** Phase II Intertidal Sample Location

Subtidal Sampling Location - Survey

• AECOM

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**The Society COMF** \*

EUNIS Biotope - Symbol Colour

A5.233 (SS.SSa.IFiSa.NcirBat) –<br>Wephtys cirrosa and Bathyporeia spp. in in A5.233 (SS.SSa.IFiSa.NcirBat) – *infralittoral sand'* 

> A5.242 (SS.SSa.IMuSa.FfabMag) – 'Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand'



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## **Sediment Composition**

The major sediment fractions at each OWF benthic grab station are presented in Figure 14D-15. The particle size analysis (PSA) data has been summarised and classified as per the Folk (1954) classification system (as described in Table F-1). There was little variation between the OWF stations, all being dominated by a high content of sandy sediments  $(63 \mu m - 2 \mu m)$ , with a generally low mud content (sediment <63 µm). Only station 21C had a sediment composition containing gravel (sediment ≥2 mm), representing 11.2% of the total sediment fraction. Overall, sand represented the highest sediment fraction across all stations (>90%), excluding station 21C (sand = 75.8%). The classification of most stations was 'sand', whilst station 21C was classified as 'gravelly muddy sand'. This conforms with the PSA results from other areas in the Tees Bay sampled as part of NZT subtidal benthic survey in 2019.



### **Figure 14D-15: Major sediment fractions (%) at each OWF grab sampling station considered**



### **Table F-3: Summarised OWF PSA data as classified by Folk (1954)**





## **Macrobenthic Communities**

Across all OWF benthic grab stations (not just those considered in this appendix), a total of 114 species were recorded, with Chaetozone cf. christiei and Magelona johnsti being the most commonly encountered species recorded. For the OWF benthic grab stations considered within this appendix, the average abundance recorded was 517.5 individuals/m2. The key species characterising each of these stations and contributing to similarity in infaunal multivariate cluster groups is outlined below.

The species richness (total number of species, S) and diversity (Shannon diversity index, H') at each OWF benthic grab station is presented in Figure 14D-16. Species richness ranged from 4 to 34 species, whilst species diversity ranged from H' = 1.034 to H' = 2.945. This was comparable to the range of species richness and diversity recorded during the Teesside Net Zero subtidal benthic surveys  $(S = 8$  to  $S = 37$ ; H' = 1.275 to H' =  $2.854$ ).





**Figure 14D-16: Species richness (S) and Shannon diversity index (H') recorded at each OWF benthic grab station considered within this appendix**

## **Priority Species and INNS**

The OWF benthic grab surveys recorded the presence of a number of individuals and colonies of Sabellaria spinulosa. This species forms biogenic reefs which is an Annex 1 habitat under the Habitats Directive, as well as being a habitat of principal importance listed under Section 41 of the NERC Act 2006. Of the OWF benthic grab stations considered within this appendix, Sabellaria spinulosa was recorded at station 21C only, with a total of 25 individuals. Overall, the results of the OWF benthic surveys concluded that the abundance of Sabellaria spinulosa was not great enough to represent biogenic reef. No other species of conservation importance were found during the OWF benthic survey; all species were considered common to the Teesside area and in UK waters.

### **Biotope Classifications**

Multivariate analysis of the OWF benthic grab stations was undertaken by Entec UK Ltd to determine the clustering of stations with a similar community composition, and to assign different biotope classifications.

Five discrete groups  $(A - E)$  were identified using cluster analysis and a SIMPROF test. Of these, groups A and B were considered as two distinct clusters, representing the majority of the grab samples. Groups C, D, and E correspond to three grab samples and do not include the stations considered within this appendix. SIMPER analysis was used to identify the species which contribute to within group similarity, and how these characterise each group. The results of this analysis11, including which stations (considered in this appendix) comprise each group, is presented in Table F-2. Nephtys cirrosa contributed the highest to the within group similarity of Group A, representing 47.04%. In Group B, both Chaetozone cf. christiei and Magelona johnsti accounted for the highest within group similarity, representing 13.00% and 11.80%, respectively.

<sup>&</sup>lt;sup>11</sup> The SIMPER analysis was undertaken for all OWF grab sampling stations, not just those considered within this appendix.



#### **Table F-2. OWF infaunal multivariate cluster groups and the results of the SIMPER analysis\***

\*top four species contributing to similarity presented

Each OWF infaunal multivariate cluster group was assigned a biotope outlined within the EUNIS habitat classification system (EEA, 2012), based on the composition of the species assemblage at each station and the composition of the substrate. Each biotope is based on codes outlined within the EUNIS habitat classification system (EEA, 2012). A habitat classification map of each station is presented in Figure 14D-14.

Group A was classified as 'Nephtys cirrosa and Bathyporeia spp. in infralittoral sand' (A5.233; SS.SSa.IFiSa.NcirBat), which is synonymous with sediment that has a high content of sand, with little to no fractions of mud ('infralittoral fine sand'). The stations comprising group A (such as 18 and 19) were found in the shallow inshore area which is characterised by moderate to high exposure and sediments possessing a low clay/silt content, characteristic of this biotope. The amphipod Bathyporeia sp. and polychaete Nephtys cirrosa are typical of this biotope and dominated the abundance of these stations.

In contrast, group B was classified as 'Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand' (A5.242; SS.SSa.IMuSa.FfabMag). This biotope is typically found in less exposed areas compared to the biotope A5.233, 'extending from the extreme lower shore down to more stable circalittoral zone at about 15-20 m' (EEA, 2019). The stations of group B were located in most cases, in slightly deeper waters and were less exposed, exhibiting a higher percentage of silt/clay. Due to the higher content of mud for this biotope, a greater dominance of venerid bivalves is expected.

The two biotopes identified (A5.233 and A5.242) qualify as habitats of principal importance listed under Section 41 of the NERC Act 2006 and belong to the habitat type, 'subtidal sands and gravels'. These are also representative of the Annex I habitat 'sandbanks slightly covered by sea water all the time'. However, these habitats are not a qualifying feature of any nearby designated site.

## **Discussion**

The sediment content of the 2019 Teesside Net Zero subtidal benthic stations in Tees Bay, consisted of predominantly sand, with a generally low mud and gravel content.
The classification of these stations was 'slightly gravelly sand', 'slightly gravelly muddy sand', and 'sand'. This conforms with the high content of sand recorded in the additional 16 OWF benthic grab samples considered within this appendix.

The Teesside Net Zero stations in Tees Bay were classified as either the biotope 'Nephtys cirrosa and Bathyporeia spp. in infralittoral sand' (A5.233; SS.SSa.IFiSa.NcirBat) or 'Fabulina fabula and Magelona mirabilis with venerid bivalves and amphipods in infralittoral compacted fine muddy sand' (A5.242; SS.SSa.IMuSa.FfabMag). In general, the stations in the shallow inshore area, where the level of exposure is considered to be greater (apparent from the lower sediment content of mud), were determined to be the biotope A5.223. The stations located in slightly deeper waters, where the sediment content of mud was higher and as such the number of venerid bivalves were also, were classified as A5.242. These two biotopes were also recorded at the OWF benthic grab stations considered within this appendix, demonstrating the same association between water depth gradients and mud gradients from the shore and the biotope assigned (see Figure 14D-14). It was noted in the OWF benthic survey report that, although small scale spatial variations between grabs were recorded, 'in terms of the specific macro-faunal assemblage', these variations were not sufficient to change the biotope classifications (Entec UK Ltd, 2011).